# OPTIMUM NO<sub>2</sub>/NO<sub>x</sub> RATIO FOR EFFICIENT SELECTIVE CATALYTIC REDUCTION<sup>1</sup>

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A new method of using ozone generator to improve selective catalytic reduction of nitrogen oxides (NO<sub>x</sub>) was investigated. In most practical exhaust gases, NO<sub>x</sub> (NO plus NO<sub>2</sub>) consists mainly of NO. Injecting ozone into the exhaust gas as a means for the oxidation of NO to NO<sub>2</sub>, it was experimentally shown that the catalytic reduction of NO<sub>x</sub> was remarkably enhanced. For the present study, a dielectric barrier discharge device using ac high voltage was employed to produce ozone, which was continuously fed to the exhaust gas. The experiments were primarily concerned with the influence of the NO<sub>2</sub>/NO<sub>x</sub> ratio on the catalytic reduction of NO<sub>x</sub> under various NO<sub>2</sub>/NO<sub>x</sub> ratios were examined. The changes in the NO<sub>2</sub>/NO<sub>x</sub> ratio as a result of the ozone injection into the exhaust gas were found to largely improve the catalytic reduction of NO<sub>x</sub>, especially at low temperatures.

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#### 1 Introduction

Removal of nitrogen oxides such as NO and NO<sub>2</sub>, called NO<sub>x</sub>, from a variety of exhaust gases is a significant industrial concern. Selective catalytic reduction (SCR) process is a well-known technology to remove NO<sub>x</sub> efficiently [1-5]. In the SCR process, ammonia (NH<sub>3</sub>) is generally used as the reducing agent, and NO<sub>x</sub> contained in the exhaust gas is converted into nitrogen (N<sub>2</sub>) and water (H<sub>2</sub>O) by reacting selectively with NH<sub>3</sub>. This process usually works well in a certain temperature range from 250 to 400°C, but at lower temperatures the NO<sub>x</sub> removal performance largely decreases [6, 7].

In most exhaust gases, the majority of  $NO_x$  is NO and the content of  $NO_2$  is less than 5%. A notable attempt to improve the  $NO_x$  removal performance at lower temperatures was to combine

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Fig. 1. Description of the catalytic  $NO_x$  reduction process enhanced by the ozone injection.

a nonthermal plasma process with the SCR [6, 8-10]. In this kind of combined process, the main role of the nonthermal plasma is to convert NO into  $NO_2$ . As reported in many papers [6, 7, 9, 11], the catalytic reduction of  $NO_x$  in the exhaust gas containing a mixture of NO and  $NO_2$  is faster than in the exhaust gas with the large majority of NO.

The other favorable method capable of oxidizing NO to  $NO_2$  is to inject ozone into the exhaust gas. In this work, the injection of ozone into the exhaust gas, instead of using nonthermal plasma directly, was suggested as a NO oxidation method for improving the SCR process at low temperatures. A dielectric barrier discharge (DBD) device was employed to produce ozone. Since the rate of the reaction between NO and ozone  $(NO+O_3 \rightarrow NO_2+O_2)$  is extremely fast, the ozone injection can be a preferable method to readily change the content of  $NO_2$ , i.e., the ratio of  $NO_2$  to  $NO_x$ . The present system was formed with a chamber to mix the exhaust gas and ozone, followed by a catalytic reactor, as shown in Fig. 1. In this chamber, a part of NO was oxidized to  $NO_2$  by ozone injected, prior to the exhaust gas entering the catalytic reactor. The exhaust gas treated by this way, containing the mixture of NO and  $NO_2$ , was then directed to the catalytic reactor where  $NH_3$  reacted with NO and  $NO_2$  to convert them into nitrogen and water. The details for the present investigation are described below.

#### 2 Experimental methods

In Fig. 2, the schematic diagram of the experimental apparatus consisting of a DBD device to generate ozone, a chamber to mix the exhaust gas and ozone, and a catalytic reactor is shown. The volume of the mixing chamber was  $245 \text{ cm}^3$ . The DBD device using ac high voltage has been discussed in detail previously [14]. The exhaust gas first treated by ozone in the mixing chamber



Fig. 2. Schematic diagram of the experimental apparatus.

entered the catalytic reactor in which a commercially available monolithic  $V_2O_5$ -WO<sub>3</sub>/TiO<sub>2</sub> catalyst (20 channels per square inch; apparent volume: 30 cm<sup>3</sup>) was placed. The vanadium and tungsten contents in the catalyst were 1.6 and 7.3 wt.%, respectively, and the surface area determined by BET (Brunauer-Emmett-Teller) adsorption isotherm was 60 m<sup>2</sup>/g. In order to adjust the reaction temperature to a desired value, the mixing chamber and the catalytic reactor were installed in an oven.

The simulated exhaust gas was prepared by mixing air-like gas (nitrogen plus oxygen) and a small amount of concentrated NO gas (5.0%(v/v)) balanced with N<sub>2</sub>). The unit %(v/v) is the percent concentration of a substance on the basis of volume, i.e., 100 times the volume of the substance of interest divided by total volume. The flow rate of the concentrated NO gas was controlled by a mass flow controller (MFC, Model 1179A, MKS Instruments, Inc., USA). In the air/NO mixture prepared as mentioned above, a small portion of NO (about 4%) was naturally oxidized to  $NO_2$  due to the presence of oxygen. The oxygen source for the ozone generator was also air. Typical inlet concentration of  $NO_x$  (NO+NO<sub>2</sub>) was 300 ppm (parts per million, volumetric; 1 ppm corresponds to  $4.1 \times 10^{-5}$  mol/m<sup>3</sup> at 298 K and 1.0 atm), and it was varied from 200 to 500 ppm. For the catalytic reduction, ammonia (NH<sub>3</sub>) was used as the reducing agent. The ratio of  $NH_3$  to  $NO_x$  concentration was typically 1.0, and it was varied from 0.5 to 1.0. The flow rate of the simulated exhaust gas was typically  $8.33 \times 10^{-5}$  m<sup>3</sup>/s (5 L/min) on the basis of room temperature. The experiments were conducted at temperatures in the range of 90 to  $230^{\circ}$ C. The NO<sub>2</sub>/NO<sub>x</sub> ratio at the inlet of the catalytic reactor was adjusted by changing the amount of ozone produced in the DBD device, i.e., the higher ac voltage applied to the DBD device was, the more ozone was produced, thereby converting more NO into  $NO_2$  in the mixing chamber placed in front of the catalytic reactor. In an experiment performed without oxygen, the simulated exhaust gas was prepared by mixing nitrogen, concentrated NO gas (5.0%(v/v))balanced with  $N_2$ ) and concentrated  $NO_2$  gas (5.0%(v/v) balanced with  $N_2$ ).

The concentrations of NO and  $NO_2$  were analyzed by a chemiluminescence  $NO-NO_2-NO_x$  analyzer (Model 42C, Thermo Environmental Instruments, Inc., USA). This kind of  $NO-NO_2-NO_x$  analyzer has a problem in measuring  $NO_2$  concentration when ammonia is present, and

thus, a Fourier Transform Infrared (FTIR) spectrophotometer (Model 1600, Perkin-Elmer, Inc., USA) equipped with a 2.4 m permanently aligned long path gas cell (Pike Technologies, Inc., USA) was utilized for the analysis of  $NO_2$  concentration. The concentration of ozone was measured by a portable gas analyzer (Porta Sens II, Analytical Technology, Inc., USA).

#### 3 Results and discussion

#### 3.1 Influence of the inlet $NO_2/NO_x$ ratio on the catalytic reduction

As described in the preceding paper [14], the injection of ozone into the exhaust gas does not lower the  $NO_x$  level (NO+NO<sub>2</sub>), simply converting NO into  $NO_2$ . But, the change in the  $NO_2/NO_x$  ratio can significantly affect the performance of the catalytic reactor. This section deals with the effect of the  $NO_2/NO_x$  ratio on the catalytic reduction of  $NO_x$ .

In the typical SCR process using  $NH_3$  as the reducing agent, the NO contained in the exhaust gas is reduced to  $N_2$  as follows:

$$4\text{NO} + 4\text{NH}_3 + \text{O}_2 \rightarrow 4\text{N}_2 + 6\text{H}_2\text{O} \tag{1}$$

Since most practical exhaust gases are made up mainly of NO, reaction (1) is the major one when ozone is not injected.

If the exhaust gas contains a considerable amount of  $NO_2$  by the ozone injection, the reaction as below rapidly occurs:

$$NO+NO_2+2NH_3 \rightarrow 2N_2+3H_2O \tag{2}$$

Besides, the following reactions can also reduce  $NO_x$  at low temperatures:

$$6NO_2 + 8NH_3 \rightarrow 7N_2 + 12H_2O \tag{3}$$

$$2NO_2 + 2NH_3 \rightarrow N_2 + H_2O + NH_4NO_3(s) \tag{4}$$

In a certain range of  $NH_3$  concentration, it does not affect the reaction rate, i.e., the reaction order of  $NH_3$  is zero, while the reaction orders of NO and  $NO_2$  are reported to be one [1, 6, 12, 13].

As an example to elucidate the behavior of the catalytic reduction according to the change in the  $NO_2/NO_x$  ratio at the inlet of the catalytic reactor, Fig. 3 shows the variations of NO,  $NO_2$ and  $NO_x$  concentrations at the outlet of the catalytic reactor when the temperature was 150°C. As the  $NO_2/NO_x$  ratio at the inlet of the catalytic reactor increased by the ozone injection, the concentration of NO at the outlet decreased while the concentration of  $NO_2$  increased. The concentration of  $NO_x$  (NO plus  $NO_2$ ) showed a minimum when the  $NO_2/NO_x$  ratio was about 0.5. From the results in Fig. 3, it is obvious that the  $NO_2/NO_x$  ratio at the inlet of the catalytic reactor can influence the  $NO_x$  removal and there exists an optimum inlet  $NO_2/NO_x$  ratio for efficient catalytic  $NO_x$  removal.

# 3.2 Effect of the temperature of gaseous mixture on the $NO_x$ removal efficiency

The temperature dependence of the catalytic reduction of  $NO_x$  is shown in Fig. 4. In this figure, the  $NO_x$  removal efficiency (Re%) was calculated as follows:

$$Re\% = 100 \times \frac{[NO_x]_0 - [NO_x]_f}{[NO_x]_0}$$
(5)



Fig. 3. Concentration variations of NO, NO<sub>2</sub> and NO<sub>x</sub> at the outlet of the catalytic reactor as a function of the inlet  $NO_2/NO_x$  ratio (reaction temperature: 150°C; initial NO<sub>x</sub>: 300 ppm; ammonia: 300 ppm; oxygen: 20%(v/v); water vapor: 0%(v/v)).

In Eq. (5), the subscripts 0 and f stand for the inlet and outlet of the reactor system, respectively. From 150 to 230°C, there was an optimal inlet  $NO_2/NO_x$  ratio at each temperature, leading to a maximum  $NO_x$  removal efficiency. Without the ozone injection, the  $NO_x$  removal efficiencies obtained at 150, 170, 200 and 230°C were 51, 67, 76 and 93%, respectively. When the  $NO_2/NO_x$  ratio at the inlet of the catalytic reactor was increased to 0.5 by the ozone injection, the respective  $NO_x$  removal efficiencies at the corresponding temperatures increased to 80, 86, 90 and 97%. As mentioned above, the catalytic reduction of  $NO_x$  via reaction (2) is faster than that via reaction (1). As well, reaction (2) implies that the content of  $NO_2$  should be equal to that of NO. That is why maximum  $NO_x$  removal efficiency occurred when the  $NO_2/NO_x$  ratio was around 0.5. On the other hand, at temperatures lower than 120°C, the  $NO_x$  removal efficiency was kept increasing with the inlet  $NO_2$  content. This result means that reactions (3) and (4) as well as reaction (2) played an important role in the  $NO_x$  removal in the low temperature range.

Meanwhile, the dependence of the  $NO_x$  removal efficiency on the reaction temperature was very large when ozone was not injected into the exhaust gas (inlet  $NO_2/NO_x$  ratio: 0.04). However, when the inlet  $NO_2/NO_x$  ratio was increased to 0.5 by the ozone injection,  $NO_x$  removal efficiency was largely enhanced, and also the temperature dependence of the  $NO_x$  removal efficiency became smaller. This small temperature dependence of the  $NO_x$  removal efficiency in the presence of the ozone injection is because the activation energy of reaction (2) is much less than that of reaction (1) [6].



Fig. 4. Temperature dependence of the catalytic reduction of  $NO_x$  at various inlet  $NO_2/NO_x$  ratios (initial  $NO_x$ : 300 ppm; ammonia: 300 ppm; oxygen: 20%(v/v); water vapor: 0%(v/v)).

# **3.3** Effect of the initial NO<sub>x</sub> concentration on the NO<sub>x</sub> removal efficiency

The effect of the initial  $NO_x$  concentration on the  $NO_x$  removal efficiency is given in Fig. 5. The initial  $NO_x$  concentration was changed from 200 to 500 ppm, fixing the ratio of  $NH_3$  to the initial  $NO_x$  to 1.0. As can be seen, the  $NO_x$  removal efficiency showed little dependence on its initial concentration over the range of 200 to 500 ppm. The fact that the  $NO_x$  removal efficiency is not affected by the initial concentration can be a merit from the practical application point of view. Compared to the case of NO-rich exhaust gas (without the ozone injection), the  $NO_x$  removal efficiency was improved by 20% when the inlet  $NO_2/NO_x$  ratio was increased to 0.5 by the ozone injection.

The main  $NO_x$  removal pathway in the absence of the ozone injection is reaction (1), which is a first order reaction in terms of NO concentration, as mentioned above. The first order reaction means that the  $NO_x$  removal efficiency does not depend on the concentration. That is why the  $NO_x$  removal efficiency without the ozone injection was nearly independent on the initial concentration. On the other hand, when NO and  $NO_2$  coexists by the ozone injection,  $NO_x$  is removed by reactions (1)-(4). Among them, reaction (2) playing important role in the removal of  $NO_x$  is a second order reaction in terms of both NO and  $NO_2$ . When the reaction order is two, the removal efficiency should depend on the concentration. But, the  $NO_x$  removal efficiency in the presence of the ozone injection was also not relevant to the initial concentration. To explain this result, more detailed reaction mechanism should be considered. According to the literature [15], the first step of reaction (2) is the adsorption of  $NO_2$  on the catalyst surface, and



Fig. 5. Effect of the initial  $NO_x$  concentration on the  $NO_x$  removal efficiency (reaction temperature: 170°C;  $NH_3$ /initial  $NO_x$ : 1.0; oxygen: 20%(v/v); water vapor: 0%(v/v)).

then the adsorbed NO<sub>2</sub> reacts with NO in the gas phase to form N<sub>2</sub>O<sub>3</sub>. The N<sub>2</sub>O<sub>3</sub> fixed on the catalyst surface is reduced to N<sub>2</sub> and H<sub>2</sub>O through the reaction with the adsorbed NH<sub>3</sub>. The formation of N<sub>2</sub>O<sub>3</sub> and the reaction of N<sub>2</sub>O<sub>3</sub> with NH<sub>3</sub> are fast whereas the adsorption of NO<sub>2</sub> on the catalyst surface is very slow. Since the adsorption of NO<sub>2</sub> that is the rate-determining step is a first order reaction in terms of NO<sub>2</sub> concentration, the NO<sub>x</sub> removal efficiency in the presence of the ozone injection was almost independent on the initial concentration.

#### **3.4** Effect of the oxygen content on the NO<sub>x</sub> removal efficiency

The effect of the oxygen content in the exhaust gas on the  $NO_x$  removal efficiency is presented in Fig. 6. In the range from 4 to 20%(v/v), the oxygen content did not largely affect the  $NO_x$ removal efficiency. Particularly, the  $NO_x$  removal efficiency at the inlet  $NO_2/NO_x$  ratio of 0.5 was similar in this range of the oxygen content. Since the oxygen content from 4 to 20%(v/v)is excessive, compared to the initial  $NO_x$  concentration of 300 ppm, the weak influence on the  $NO_x$  removal is reasonable. Meanwhile, when oxygen was not present, lower  $NO_x$  removal efficiencies were observed. For reaction (1), oxygen is necessary. Without oxygen, this reaction occurs slowly, resulting in low  $NO_x$  removal efficiency. When NO and  $NO_2$  coexists by the ozone injection, reaction (2) that does not require oxygen can significantly contribute to the removal of  $NO_x$ . Thus, high  $NO_x$  removal efficiency can be achieved even in the absence of oxygen.



Fig. 6. Effect of the oxygen content on the  $NO_x$  removal efficiency (reaction temperature: 170°C; initial  $NO_x$ : 300 ppm; ammonia: 300 ppm; water vapor: 0%(v/v)).

# **3.5** Effect of the water vapor content on the $NO_x$ removal efficiency

Real exhaust gases from hydrocarbon combustions contain considerable amounts of water vapor. So as to examine the effect of the water vapor content on the catalytic reduction of  $NO_x$ , it was varied between 0 and 5%(v/v) with the other conditions kept constant. Since water molecules can occupy some of the active sites of the catalyst, they may inhibit the catalytic reduction of  $NO_x$ . From Fig. 7, it is apparent that the increase in the inlet  $NO_2/NO_x$  ratio by the ozone injection resulted in a remarkable improvement in the catalytic reduction of  $NO_x$  regardless of the presence of the water vapor. When the water vapor content increased from 0 to 1.5%(v/v), slight decreases in the  $NO_x$  removal efficiency were observed both with and without the ozone injection. On the contrary, further increases in the water vapor content up to 5% (v/v) did not nearly affect the catalytic reduction. It is also reported in the literature that water vapor above 1.0%(v/v)exhibits similar inhibiting action on the catalytic reduction of  $NO_x$  [16]. This phenomenon may be explained by the high adsorption energy of ammonia. The catalytic reduction of  $NO_x$  gets started by the adsorption of ammonia, which reacts with nitrogen oxides molecule in the next step. Thus, the adsorption of ammonia plays an important role in the catalytic reduction process. Generally, the adsorption energy of ammonia is higher than that of water vapor, i.e., the water molecule adsorbed on the catalyst surface can be easily replaced by ammonia.



Fig. 7. Effect of the water vapor content on the  $NO_x$  removal efficiency (reaction temperature: 170°C; initial  $NO_x$ : 300 ppm; ammonia: 300 ppm; oxygen: 20%(v/v)).

# 3.6 Effect of the ammonia concentration on the NO<sub>x</sub> removal efficiency

Fig. 8 presents the effect of the ammonia concentration used as the reducing agent on the  $NO_x$  removal efficiency. In order to examine the effect of the ammonia concentration, the ratio of  $NH_3$  to the initial  $NO_x$  concentration was changed from 0.5 to 1.0 with the other variables fixed. One general trend is that the  $NO_x$  removal efficiency increased with the amount of  $NH_3$  added, which indicates that the reducing agent is necessary in the catalytic  $NO_x$  removal process. In the same manner with the above results, best  $NO_x$  removal efficiency was observed at the  $NO_2/NO_x$  ratio of 0.5, i.e., when the concentration of NO was equal to that of  $NO_2$ . This result emphasizes again that the amount of ozone injected into the exhaust gas should be adjusted to give equimolar NO and  $NO_2$  mixture.

#### 3.7 Comparison between the present process and the plasma-catalysis hybrid process

A comparison between the nonthermal plasma-catalysis hybrid process and the present process was made as in Fig. 9. So as to compare the results on an identical basis, the data for the  $NO_x$  removal efficiencies were presented with respect to the energy density applied to the nonthermal plasma reactor or the ozone generator. In case of the nonthermal plasma-catalysis process using dielectric barrier discharge [9], both ethylene and ammonia were used because the rate of NO oxidation was very slow without hydrocarbon additive. As can be seen, the  $NO_x$  removal efficiency obtained in the present process was similar to that obtained in the nonthermal plasma-catalysis



Fig. 8. Effect of the ammonia concentration on the  $NO_x$  removal efficiency (reaction temperature: 170°C; initial  $NO_x$ : 300 ppm; oxygen: 20%(v/v); water vapor: 0%(v/v)).

hybrid process when the energy density was below 25 J/L. However, at higher energy densities the nonthermal plasma-catalysis hybrid process showed better  $NO_x$  removal performance than the present process. As reported in the previous study [6, 9], the nonthermal plasma can remove some  $NO_x$  although the main role of the nonthermal plasma is to oxidize NO to  $NO_2$ . In Fig. 9, the better  $NO_x$  removal performance of the plasma-catalysis hybrid process may be attributed to the removal of  $NO_x$  by the action of the nonthermal plasma. But, the formation of unwanted byproducts from hydrocarbon additive when using the nonthermal plasma should be resolved. In this context, the present ozone injection method producing no byproducts may be more promising for practical uses.

### 4 Concluding remarks

The application of ozone injection method to the catalytic reduction of nitrogen oxides  $(NO_x)$  was investigated. The addition of ozone to the exhaust gas increased the  $NO_2/NO_x$  ratio as a result of the oxidation of NO to  $NO_2$ , leading to the enhancement in the catalytic reduction of  $NO_x$ . Main conclusions drawn in this study can be summarized as follows. Above 150°C, there was an optimum inlet  $NO_2/NO_x$  ratio for the efficient catalytic reduction of  $NO_x$ , which was found to be around 0.5. On the other hand, at lower temperatures less than 120°C, the  $NO_x$  removal efficiency in the catalytic reactor tended to keep increasing with the inlet  $NO_2/NO_x$  ratio. The  $NO_x$  removal efficiency showed large dependence on the reaction temperature, but



Fig. 9. Comparison of the  $NO_x$  removal efficiency between the present hybrid process and the nonthermal plasma-catalysis hybrid process (reaction temperature: 150°C; initial  $NO_x$ : 300 ppm; ammonia: 300 ppm; oxygen: 20%(v/v); water vapor: 0%(v/v)).

the relationship between the  $NO_x$  removal efficiency and its initial concentration in the range of 200 to 500 ppm was hardly observed. Besides, the oxygen content from 4 to 20%(v/v) and the water vapor content from 1.5 to 5.0%(v/v) did not largely affect the  $NO_x$  removal efficiency. The ozone injection method was very efficient to improve the catalytic reduction of  $NO_x$  in a wide range of temperatures, and the  $NO_x$  removal efficiency was comparable to the nonthermal plasma-catalysis hybrid process.

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